

UNIVERSIDADE FEDERAL DE SÃO CARLOS  
CENTRO DE CIÊNCIAS BIOLÓGICAS E DA SAÚDE  
PROGRAMA DE PÓS-GRADUAÇÃO EM ECOLOGIA E RECURSOS NATURAIS

**CLARISSA DE OLIVEIRA SANTOS**

EFFECTS OF LOCAL AND LANDSCAPE ENVIRONMENTAL VARIABLES ON AVIFAUNA  
ABUNDANCE IN FOREST FRAGMENTS OF ATLANTIC FOREST IN SOUTHEASTERN BRAZIL

Dissertação apresentada ao Programa  
de Pós-Graduação em Ecologia e  
Recursos Naturais, para obtenção  
do título de Mestre em Ecologia  
e Recursos Naturais.

São Carlos  
2018

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Effects of local and landscape environmental variables on avifauna abundance in forest fragments of Atlantic  
Forest in southeastern Brazil

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Co-orientação: Prof. Dr. Rogério Hargunt Toppa

São Carlos

2018



## UNIVERSIDADE FEDERAL DE SÃO CARLOS

Centro de Ciências Biológicas e da Saúde  
Programa de Pós-Graduação em Ecologia e Recursos Naturais

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Prof. Dr. Augusto João Piratelli

## Agradecimentos

Ao Prof. Dr. Augusto João Piratelli, pela orientação e apoio nesse trabalho, pela paciência e gentileza diante a todas minhas dificuldades.

Ao Prof. Dr Rogério Hargunt Toppa, pela co-orientação e coleta de dados relacionados a paisagem.

À Prof Dr. Fátima Piña-Rodrigues e ao LASEM, pelo auxílio e empréstimo de equipamentos.

Ao Prof. Dr. Marurício Cetra pelas contribuições nas análises.

À Secretaria do Meio Ambiente de Araçoiaba da Serra pelo auxílio.

À Monica Alcarás de Góes e ao Jorge Vieira de Góes pelo contato com os proprietários.

A todas as pessoas maravilhosas que me ajudaram nas coletas, “Miudo”, “Aska”, Mariana, “Oti” e Mirelinha linda.

Às pessoas que me ajudaram com os áudios, Ana Luiza, “Filcu” e “Boga”.

Os proprietários e funcionários das fazendas nas quais os dados foram coletados.

Aos meus colegas do LECO, que me receberam tão bem nesses dois anos, Dani, Maristela, Paula, Rafaela, “Mascote”, Paulo, Bianca, e principalmente ao Lucas e à Raíssa, que me ajudaram nas coletas e foram ótimos companheiros de mestrado.

Ao Luís por disponibilizar seu carro para as coletas.

Aos membros da banca de qualificação, Dr. Carlos Candia-Gallardo (“Kiwi”), Prof dr. Eduardo Moschini e Prof. Dr. Marco Batallha, pelas contribuições no trabalho.

Às minhas irmãs, de todas as gerações, da República DiLadinho, por sempre me apoiarem.

Aos queridos da República Etanóis que me acolheram da melhor maneira possível durante as provas do mestrado.

À galera da República GreenHouse, Livia, “Tiziu”, “Nardoni”, Bruna, Thais, Fábio, Helena e “Testa”, por serem minha família em São Carlos.

À Mariana por dividir sua casa comigo, e a toda família Bergamo Caraméz que me acolhe desde 2005, em São Paulo, Botucatu e agora Sorocaba.

À galera da Casita, “Guizbi”, “Tê”, “Guigo”, “Búfalo” e “Ju” pelo abrigo em São Paulo, e por todo apoio para que eu conseguisse concluir essa etapa.

Ao meu “Lhofi”, namos lindo, por todo apoio e carinho, por acreditar em mim e pela vida que compartilhamos

Ao Arthurzinho, meu parceiro das risadas, das danças, das tretas e da vida toda.

Ao Benedito e a Yara, por todo amor, carinho, ensinamentos e investimentos nesses 26 anos.

À Coordenação de Aperfeiçoamento de Pessoal de Nível Superior (CAPES) (processo nº 1586584) pelo apoio financeiro.

E por fim, gostaria de agradecer imensamente ao Cícero Branco, que foi essencial para que este trabalho se concretizasse, muito obrigada por compartilhar sua experiência, e se dispor a me ajudar na maioria das coletas.

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## RESUMO GERAL

A conversão de florestas neotropicais em terras agrícolas tem como consequências alterações na estrutura interna e pressões externas sobre remanescentes florestais, que afetam a abundância de aves em diferentes escalas espaciais. Neste estudo, verificamos a contribuição relativa de três escalas espaciais (local, imediato e de paisagem) sobre a abundância de aves em fragmentos de mata atlântica no sudeste do Brasil. Nós testamos se as espécies que responderam de maneira semelhante às variáveis ambientais estão na mesma guilda de forrageamento. Realizamos este estudo através de 44 pontos fixos em 18 fragmentos, na zona de amortecimento de uma unidade de conservação. Foram amostradas 15 variáveis ambientais em três escalas espaciais; seis locais específicos, três nos arredores imediatos (raio de amortecimento de 100m) e seis na paisagem). Foram registrados 1724 indivíduos de 108 espécies entre abril de 2017 e fevereiro de 2018. Classificamos as espécies amostradas em 15 grupos de forrageamento. As frações únicas de local e de paisagem afetaram a abundância de aves, mas os arredores imediatos e exclusivos não influenciaram significativamente essa variável. Identificamos um padrão relacionado aos efeitos da escala local apenas para aves insetívoras, mas nenhum padrão para as outras guildas de forrageamento associadas a quaisquer outras escalas. Nosso estudo pode subsidiar práticas de manejo nos locais avaliados, incrementando principalmente a densidade do sub-bosque e a altura do dossel remanescente das florestas nativas.

**Palavras-chave:** aves, fragmentos florestais, grupos tróficos, zona de amortecimento

## **Artigo Produzido**

Este estudo gerou um artigo científico, nomeado “Effects of local and landscape environmental variables on avifauna abundance in forest fragments of the Atlantic Forest in southeastern Brazil”, que foi submetido à revista *Emu Austral Ornithology*, submissão que é solicitada na norma regimental atual do Curso de Pós-Graduação em Ecologia e Recursos Naturais da Universidade Federal de São Carlos. Este trabalho está formatado segundo as normas exigidas pela revista.



**Effects of local and landscape environmental variables on avifauna abundance in forest fragments of the Atlantic Forest in southeastern Brazil**

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# **Effects of local and landscape environmental variables on avifauna abundance in forest fragments of the Atlantic Forest in southeastern Brazil**

## **ABSTRACT**

Pristine Neotropical forests have been converted into agricultural lands. Shifts in the internal structure and external pressures on forest remnants affect bird assemblages that require specific resources for survival at different spatial scales. In this study, we verified the relative contribution of three spatial scales (local, immediate surroundings and landscape) on the abundance of birds in Atlantic forest fragments in south-eastern Brazil. We tested whether species that responded in a similar way to environmental variables are in the same foraging guild. We carried out this study through 44 fixed points in 18 fragments, in the buffer zone of a protected area. We sampled 15 environmental variables in three spatial scales; six site-specific, three at the immediate surrounds (buffer radius of 100m), and six at the landscape). We recorded 1724 individuals from 108 species between April 2017 and February 2018. We classified sampled species in 15 foraging groups. The unique fractions of local and landscape have affected bird abundance, but the exclusive immediate surroundings have not significantly influenced this variable. We have identified a pattern related to the effects of local scale only for insectivorous birds, but no pattern for the other foraging guilds linked to any other scales. Our study may support management practices in the studied sites, practices mainly improving the understory density and canopy height of the remaining patches of native forest.

**Keywords:** Birds, buffer zone, forest patches, trophic groups

## ***Introduction***

Forest fragmentation resultant of anthropogenic activities has been cited as one of the major causes of global biodiversity decline (D'Eon *et al.* 2002; Gardner *et al.* 2009; Barlow *et al.* 2016). The habitat loss exposes organisms to several stresses, which occur at different spatial

scales (Blaustein *et al.* 2011; Chase and Knight 2013; Bartlett *et al.* 2016) and may vary in intensity among taxa and ecological groups (Leibold *et al.* 2004; Ewers and Didham 2006). Thus, recent research has been increasingly investigating the effects of fragmentation at different scales on several groups of organisms (Kennedy *et al.* 2013; Sarthou *et al.* 2014; Lafage and Pétilion 2016; Morante-Filho *et al.* 2016).one of the major causes of global biodiversity decline (D'Eon *et al.* 2002; Gardner *et al.* 2009; Barlow *et al.* 2016). The habitat loss exposes organisms to several stresses, which occur at different spatial scales (Blaustein *et al.* 2011; Chase and Knight 2013; Bartlett *et al.* 2016) and may vary in intensity among taxa and ecological groups (Leibold *et al.* 2004; Ewers and Didham 2006). Thus, recent research has been increasingly investigating the effects of fragmentation at different scales on several groups of organisms (Kennedy *et al.* 2013; Sarthou *et al.* 2014; Lafage and Pétilion 2016; Morante-Filho *et al.* 2016).

The sensitivity of bird species to fragmentation is variable and depends on several factors, such as habitat and diet specialization, population size, body size, mobility and nesting (Murphy 2012; Neuschulz *et al.* 2013; Newbold 2013; Bregman 2014; Pavlacky *et al.* 2015; Alexandrino *et al.* 2016). More specifically, the analysis of bird guilds response to habitat conditions in fragmented areas, at different spatial scales, is a recurrent approach (Pearson 1975; Altshuler 2006; Gray *et al.* 2007; Hasui *et al.* 2007; Piacentini and Varassin 2007; Boscolo and Metzger 2009; Galitsky 2015).

Guild is a term coined to refer to a set of species that explore a given resource in similar patterns (Root 1967; Blondel 2003) e.g. habitat structure, food and others. Nevertheless, the classification of species into guilds is useful since fragmentation causes changes to the habitat structure, limiting access to specific resources, such as foraging and nesting grounds (Lamanna and Martin 2016; Verga *et al.* 2017). Additionally, such categorization allows going beyond community composition and evaluating possible losses of ecosystem functions (Gray *et al.* 2007; Bovo *et al.* 2018).

However, a priori species allocation in different guilds based on *e.g.* expert knowledge and general empirical observations may bring some limitations. More specifically, species have been reported to present different responses within a given guild; therefore, generalizations at the guild level may not fit the ecological reality. For instance, the effects of fragmentation on the abundance of understory insectivorous birds vary considerably among species; imposing

limitations to determine the influence of different spatial scales on species of the same foraging guild (Banks-Leite *et al.* 2013; Edwards *et al.* 2013). A similar conflict arises for large frugivores. This foraging guild would theoretically be more susceptible to regional effects; since most species need to move across patches of native vegetation in search of fruits, which are often scarce in small fragments (Mueller *et al.* 2014; Santos-Junior *et al.* 2016). Yet, some species often included in this guild (*e.g.* *Patagioenas picazuro*; Athiê and Dias, 2010; Cruz and Piratelli, 2011; Vitorino *et al.* 2018), may have low sensitivity to anthropic alterations, being abundant even in disturbed areas (Franz *et al.* 2010; Cruz and Piratelli 2011; Alexandrino *et al.* 2013; Alexandrino *et al.* 2016). Moreover, although guilds have been identified as useful indicators, the sensitivity patterns may be context specific (Sigel *et al.* 2010, Thornton *et al.* 2011, Vetter *et al.* 2011).

Birds respond to habitat alterations by either increasing or reducing species abundance. Thus, the likelihood of detecting biodiversity responses to habitat alterations increases by monitoring such ecological variables (Mazaris *et al.* 2013; Aebischer *et al.* 2016). Ultimately, such analysis becomes a valuable tool for suggesting priorities in conservation efforts and illustrating the representativeness of protected areas (Horn and Koford 2006; Johnston 2015).

Here, we studied forest fragments of semi-deciduous seasonal Atlantic forest and tested the relative contribution of environmental variables of three spatial scales - local, immediate surrounding (100-m radius buffer) and landscape - on bird species abundance. Considering that species formally grouped in foraging guilds often vary in ecological and morphological characteristics, we expect to find no homogeneous pattern of responses in abundance within each guild.

## **Methods**

### ***Study area***

We carried out this study in an area of ~2,350ha, representing nearly 10% of the boundaries of the city of Araçoiaba da Serra, state of São Paulo, Brazil (23 ° 30 '16' 'S, 47 ° 36' 52"W) (Figure 1), which lies in a transition region between Atlantic Forest and Cerrado. The mean annual local temperature and rainfall are 20.8 ° C and 1248.5 mm, respectively (CEPAGRI, 2017). The city is part of the buffer zone of Floresta Nacional de Ipanema, a 5,069ha federal protected area

(ICMBio, 2017). Forest fragments occupy 9.4% of the city area, and the predominant vegetation is semi-deciduous seasonal forest, being 1.18% of woodlands and 8.25% of “capoeiras” (Albuquerque and Rodrigues 2000, SIFESP, 2001; IBGE 2015). Land use is characterized by pasturelands and agriculture, mainly of sugarcane crops (Sales *et al.* 2016).

### ***Avifauna***

To test the influence of local and landscape factors on avifauna, we set 44 points in 18 fragments (Figure 2). To ensure the independence of the bird data, the points were established with a minimum distance of 200m (Blake 1992; Vielliard *et al.* 2010). We conducted 50-m fixed-radius point-count surveys during 10 minutes at the 44 points twice, both in dry and wet season, in the morning and afternoon (6:00 a.m. to 9:00 a.m. and 4:30p.m. to 6:30 p.m.) (Becker *et al.* 2013, Alexandrino *et al.* 2016, Uezu and Metzger 2016). We used a Bushnell Natureview 8x42 binocular, a Sony ICD-PX440 Digital Audio Recorder, and specific literature (Ridgely *et al.* 2015) for species identification. The species nomenclature followed the classification of Handbook (HBW) and BirdLife International (2016). For each species, we calculated the index of point abundance (IPA) dividing the number of contacts (i.e. birds that were heard or seen) by the number of sampled point counts.

We classified the trophic/foraging guilds according to the diet, morphological and foraging characteristics (*e.g.* substrate and strategy of foraging). Species were split into fifteen categories according to Wilman *et al.* 2014, Del Hoyo *et al.* 2016 and Santos-Junior *et al.* 2016: foliage insectivores (Fi); omnivores (Om); understory insectivores (Ui); insectivores trunk-climbers (Ti); large frugivores (Lf); forest edge granivores (Eg); aerial insectivores (Ai); canopy insectivores (Ci); nectarivores (Ne); small frugivores (Sf); open area granivores (Ag); open area insectivores (Ia); detritivores (De); carnivores (Ca) and piscivores (Pi).

### ***Environmental variables***

We assessed the environmental condition at three spatial scales most likely to influence the avifauna (Pearson 1975; Hasiu *et al.* 2007; Piacentini and Varassin 2007). The first scale

represents the local variables, sampled in 10x10m plots around the center of each of the 44 points. Six environmental variables were assessed at each of the avifauna data collecting points, from March to August 2017. The variables were (a) richness of arboreal morphospecies; (b) number of standing dead trees; (c) average canopy height (using a laser rangefinder, by readings at the four cardinal points in the center of each plot); (d) canopy cover (by averaging four densitometer readings at the in four cardinal points in the center of the plot); (e) understory density (by counting the regenerating and shrub individuals between 50 cm and 2 m); and (f) average of the soil temperature of the four days that each point was sampled (measured with a digital thermometer in the center of the plot).

The second variable set was the immediate surrounding environment, which we sampled using a 100-m radius buffer around each point. In each buffer we calculated the forest cover area; average NDVI (normalized difference vegetation index), as a green biomass indicator (Rouse *et al.* 1973; Freitas *et al.* 2005); and the area of five classes of declivity, 1) 0° to 5°; 2) 5 to 12°; 3) 12 to 30°; 4) 30 to 45° and 5) > 45°. The last variable group was of the landscape variables, related to distance metrics. For this, we measured the distance from each point to the nearest forest patch, pond, stream and headwater, and the areas of each of these landscape features. These two last variable groups was obtained from Landsat 8 a 30-m resolution land-cover layer with 74.6% overall fuzzy accuracy in ArcGis v.10.5 (ESRI, 2016).

### ***Data analyses***

We performed a redundancy analysis (RDA) with all variables and used the Ordistep function to first, exclude the ones that were strongly correlated and second, select those that best explained the bird data variation. With the selected variables in each set, we partitioned the data variance into seven independent fractions: (a) exclusive local effects; (b) exclusive immediate surrounding; (c) exclusive landscape effects (100-m radius buffer) effects; (d) shared effects between local and immediate surrounding; (e) shared effects between landscape and immediate surrounding; (f) shared effects between local and landscape; (g) shared effects between the three set variables. Species were clustered by the *k-mean* method according to their coordinates on the

RDA axes. The distribution of the species in the RDA axes was divided into seven clusters (Fig.6), however, only species at the five clusters that presented greater variation in the RDA axes were analyzed.

In order to meet the assumptions of linearity of the RDA analysis, we standardized both the variable and bird data, using the Hellinger transformation for the later. We considered a level of significance equal to or less than 0.05. All analyzes were performed in R (R Development Core Team 2014) with the Vegan package (Oksanen et al. 2013).

## Results

We recorded 1724 bird contacts from 108 species. The most abundant species were *Basileuterus culicivorus* (IPA=4.068) and *Turdus leucomelas* (IPA=2.363) (Table S1) whereas the most abundant foraging guild were foliage Insectivores and Omnivores (Fig.3).

The variance fractions related to local, immediate surroundings (100m radius buffer) and landscape together (a+b+c+d+e+f) explained 13.7% of the variation in bird abundance (Fig. 2). The unique fractions of each set of explanatory variables, i.e. local (a) and landscape (c) were significant ( $p=0.0039$  and  $p=0.00099$ , respectively), but the exclusive immediate surroundings (b) did not significantly influence bird abundances ( $p=0.187$ ) (Table 1). The ordistep function kept the environmental variables (a) canopy height and understory density, for local scale; (b) declivity 1 ( $0-5^\circ$ ) for the immediate surroundings; and (c) nearest pond distance, for the landscape scale (Fig. 5). The first RDA axis explained 39.4% of the variance and was influenced by the understory density and the nearest pond distance. The second RDA axis explained 27.9% and was associated with canopy height and declivity between 0 and  $5^\circ$ .

The five species groups selected according to the clusters were: (a) positively related to canopy height and understory density; (b) positively related to understory density; (c) positively related to declivity between  $0-5^\circ$ ; (d) positively related to nearest pond distance; (e) negatively related to nearest pond distance (Table 3).

## Discussion

In our study, we found that landscape and local environmental characteristics have influenced the abundance of birds in semideciduous seasonal forest fragments. According to the variance partitioning, the variables of these two scales present similar pattern of influence, with the total and the unique effect of landscape characteristics (8.7% and 4.4%, respectively) showing larger influence than local scale (5.75% and 3.1%) (Fig. 4; Table 1). These results agree with previous studies, in which both spatial scales, i.e. the plot structure (here presented as local characteristics) and the landscape characteristics, influence bird richness and abundance, whereas a higher influence is observed at the landscape scale (Carrara, 2015; Galitsky 2015; Buffa et. al, 2018).

Regarding the local environmental variables, ten species were positively influenced by understory density and canopy height. Ten were insectivores with different foraging substrates. *Basileuterus culicivorus*, *Thamnophilus caeruleus* and *Vireo olivaceus* are foliage-gleaners. *Corythopsis delalandi*, *Myiothlypis flaveola* and *Tapera naevia* are understory insectivores. *Conirostrum speciosum* forages on the canopy; and *Sittasomus griseicapillus* is a trunk-insectivore (Table 3 A-B). Understory density are intimately related to insect's availability (Streby et al. 2011; Powell et al. 2015) and canopy height with patch structure and conservation (Peña-Claros 2003).

The declivity between 0-5° has positively influenced eleven species. The most abundant species were the omnivores *Turdus leucomelas* and *Pitangus sulphuratus*, and the large frugivore *Patagioenas picazuro* (Table 3C); all widely common in human-modified environments (Pereira and Brito, 2005; Franz et. al., 2010; Cruz and Piratelli, 2011; Alexandrino et. al., 2013; Alexandrino et. al. 2016). Although some of these species belong to more specialized foraging guilds (e.g., trunk-twig insectivores and nectarivores; Heyman 2010), their representatives are mostly generalists, also known to be abundant in anthropogenic environments, i.e. *Colaptes campestris* and *Coereba flaveola* (Dario 2017, Del Hoyo et al., 2018). The slope of the terrain may be related to patterns of occupation and land use in agricultural fields, since the difficulty of access to steeper areas might translate into higher deforestation of lowlands (Silva et al., 2007; et al., 2010). Thus, an inverse relationship between the declivity of the terrain and the abundance of



generalist species is expected, since these species are able to be benefited from anthropogenic disturbances (Laurance et al., 2002, Le Viol et al. 2012).

We identified a pattern among the effects of local scale on insectivorous birds, as in Bregman et al. (2015). However, this was restricted to birds' diet, not to the foraging guild (Table 3A-B). There was no pattern for foraging guilds linked to the other two scales (i.e. landscape and immediate surroundings). Our results agree with Ulrich et al. (2017), who tested a series of functional traits more liable of explaining part of the spatial segregation of birds in fragments of the Atlantic forest and found no influence related to foraging traits.

In conclusion, as a buffer zone of a protected area, our area of study still support several forest dweller species. However, its current fragmentation and conversion into intensive crop and livestock farming can jeopardize the persistence of several species, especially the more specialized in their ecological requirements. Thus, considering the trends of influence observed in our analyzed environmental variables, we propose the introduction of management practices that improve the understory density and canopy height of the remaining patches of native forest.

## **Acknowledgments**

We thank the Coordenação de Aperfeiçoamento de Pessoal de Nível Superior (CAPES), (process no. 1586584) for financial support, the owners and employees of the farms in which the study was conducted, the Secretary of the Environment of Araçoiaba da Serra by the support. We also thank Emily Hagen for English review, Carlos Candia-Gallardo for suggestions on the earlier version of this manuscript, and Cícero Branco, for help with fieldwork.

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## TABLES

Table 1. Degree of freedom (df), explained variation ( $R^2$ ), adjusted explained variation (Adj. $R^2$ ) of the three explanatory scales on the abundance of birds, in a region of the Atlantic Forest in southeastern Brazil. Fractions are as follows: exclusive effects: a= local, b= landscape c = immediate surrounding, and shared effects: d = local and landscape, e= landscape and immediate surrounding, f= local and immediate surrounding, g= local, landscape and immediate surrounding.

<b>Fractions</b>	Df	R.square	Adj.R.square
[a+d+f+g] = Local	3	0,123	0,05722
[b+d+e+g] = Immediate surrounding	2	0,10597	0,06236
[c+e+f+g] = Landsacpe	6	0,21445	0,08707
[a+b+d+e+f+g] = Local + Immediate surrounding	5	0,19837	0,0929
[a+c+d+e+f+g] =Local + Landsacpe	9	0,31203	0,12992
[b+c+d+e+f+g] = Landscape + Immediate surrounding	8	0,27244	0,10613
[a+b+c+d+e+f+g] = All	11	0,35784	0,13709
<b>Individual fractions</b>			
[a] = Exclusively Local	3	0,03096	
[b] = Exclusively Immediate surrounding	6	0,0442	
[c] = Exclusively Landscape	2	0,00717	

Table 2. Degrees of freedom (df), F statistics and p-value for the influence of effects of local, landscape and immediate surroundings on birds abundance in a region of the Atlantic Forest in southeastern Brazil. Significance levels:  $p < 0.05$ : \*0,05; \*\*0,01; \*\*\* <0,001.

	Local			
	Df	Variance	F	Pr(>F)
Model	3	0.05094	1.3244	0.00099
Residual	32	0.38304		
	Immediate surrounding			
	Df	Variance	F	Pr(>F)
Model	2	0.02732	1.1413	0.187
Residual	32	0.38304		
	Landscape			
	Df	Variance	F	Pr(>F)
Model	6	0.09512	1.3244	0.00099
Residual	32	0.38304		

Table 3. Cluster of species (a) Positively related to Canopy Height and Understorey Density; (b) Positively related to Understorey Density; (c) Positively related to Declivity 1 (0-5°); (d) Positively related to Nearest Pond Distance; (e) Negatively related to Nearest pond

	Species	Acronym	IPA	Guild
<b>a) Positively related to “Canopy Height” and “Understorey Density”</b>	<i>Basileuterus culicivorus</i>	Bcul	4,068	Fi
	<i>Thamnophilus caerulescens</i>	Tcae	1,364	Fi
	<i>Turdus albicollis</i>	Talb	0,386	Om
	<i>Corythopsis delalandi</i>	Cdel	0,818	Ui
	<i>Conirostrum speciosum</i>	Cspe	0,182	Ci
<b>b) Positively related to “Understorey Density”</b>	<i>Sittasomus griseicapillus</i>	Sgri	1,205	Ti
	<i>Vireo olivaceus</i>	Voli	1,023	Fi
	<i>Myiothlypis flaveola</i>	Mfla	1,091	Ui
	<i>Tapera naevia</i>	Tnae	0,50	Ui
	<i>Cyclarhis gujanensis</i>	Cguj	2,068	Om
<b>c) Positively related to “Declivity 1 (0-5°)”</b>	<i>Turdus leucomelas</i>	Tlcm	2,364	Om
	<i>Colaptes campestris</i>	Ccam	0,114	Ti
	<i>Patagioenas picazuro</i>	Ppic	2,25	Lf
	<i>Cyanocorax cristatellus</i>	Ccri	0,091	Om
	<i>Crotophaga ani</i>	Cani	0,182	Oi
	<i>Myiozetetes similis</i>	Msim	0,295	Om
	<i>Coereba flaveola</i>	Cfla	0,295	Ne

	<i>Pitangus sulphuratus</i>	Psul	0,864	Om
	<i>Caracara plancus</i>	Cplan	0,159	Ca
	<i>Ramphastos toco</i>	Rtoc	0,250	Lf
	<i>Hirundinea ferruginea</i>	Hfer	0,068	Ai
	<i>Turdus leucomelas</i>	Tlcm	2,364	Om
<b>d) Positively related to “Nearest Pond Distance”</b>	<i>Hylatomus lineatus</i>	Hlin	0,068	Ti
	<i>Euphonia chlorotica</i>	Echl	0,477	Sf
	<i>Phaethornis pretrei</i>	Ppre	0,182	Ne
	<i>Camptostoma obsoletum</i>	Cobs	0,455	Fi
	<i>Sicalis flaveola</i>	Sfla	0,114	Ag
	<i>Phyllomyias plumbeiceps</i>	Pplum	0,045	Ui
<b>e) Negatively related to “Nearest pond Distance</b>	<i>Dysithamnus mentalis</i>	Dmen	1,523	Fi
	<i>Lathrotriccus eulerei</i>	Leul	0,523	Ui
	<i>Leptotila verreauxi</i>	Lver	0,705	Eg
	<i>Progne tapera</i>	Ptap	0,159	Ai
	<i>Todirostrum poliocephalum</i>	Tpoli	0,091	Ci
	<i>Synallaxis ruficapilla</i>	Sruf	0,273	Fi
	<i>Piaya cayana</i>	Pcay	0,318	Om
	<i>Aramides saracura</i>	Asar	0,114	Om
	<i>Dendrocincla turdina</i>	Dtur	0,50	Ti
	<i>Myiodynastes maculatus</i>	Mmac	0,409	Om
	<i>Sclerurus scansor</i>	Ssca	0,205	Ui
	<i>Thamnophilus doliatus</i>	Tdol	0,455	Fi
	<i>Lochmias nematura</i>	Lnem	0,250	Ui
	<i>Tolmomyias sulphurescens</i>	Tsul	0,795	Fi

# FIGURES

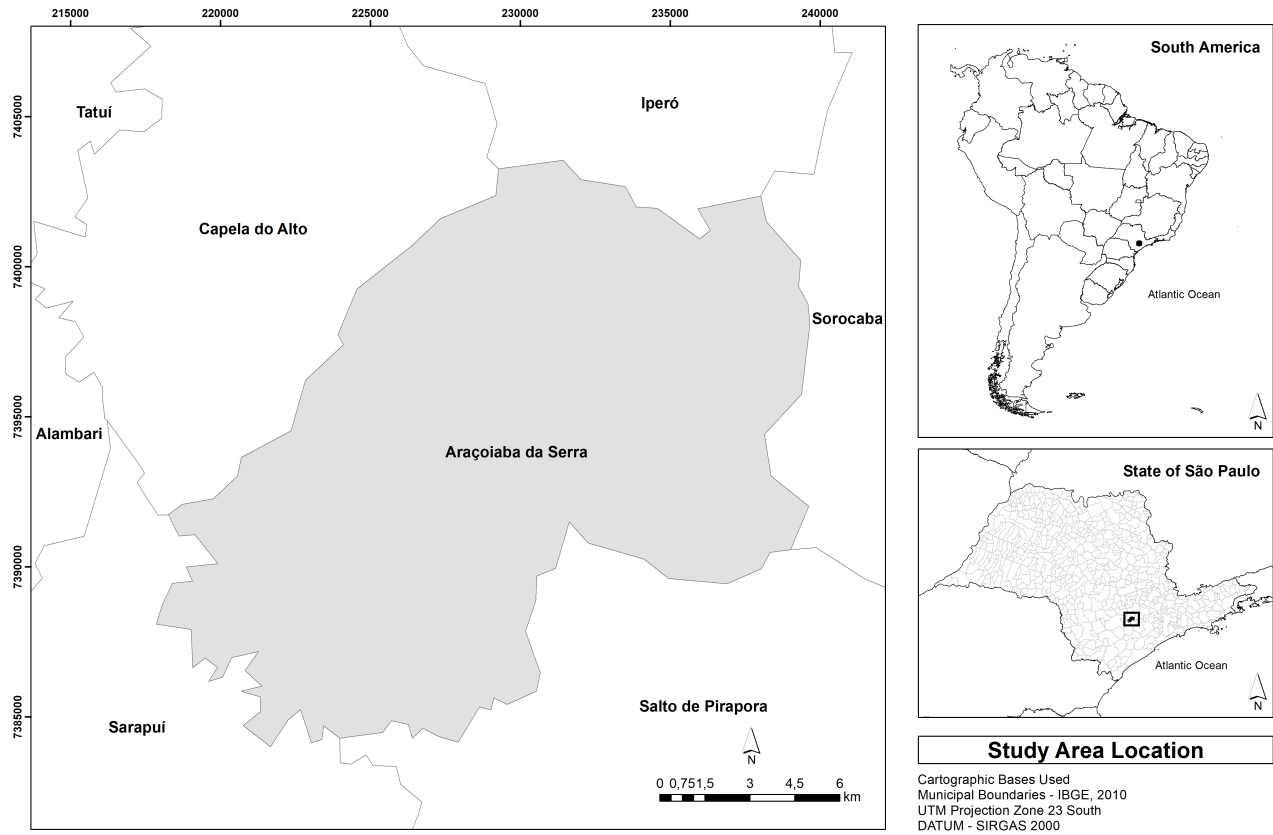


Figure 1. Araçoiaba da Serra, São Paulo, Brazil.

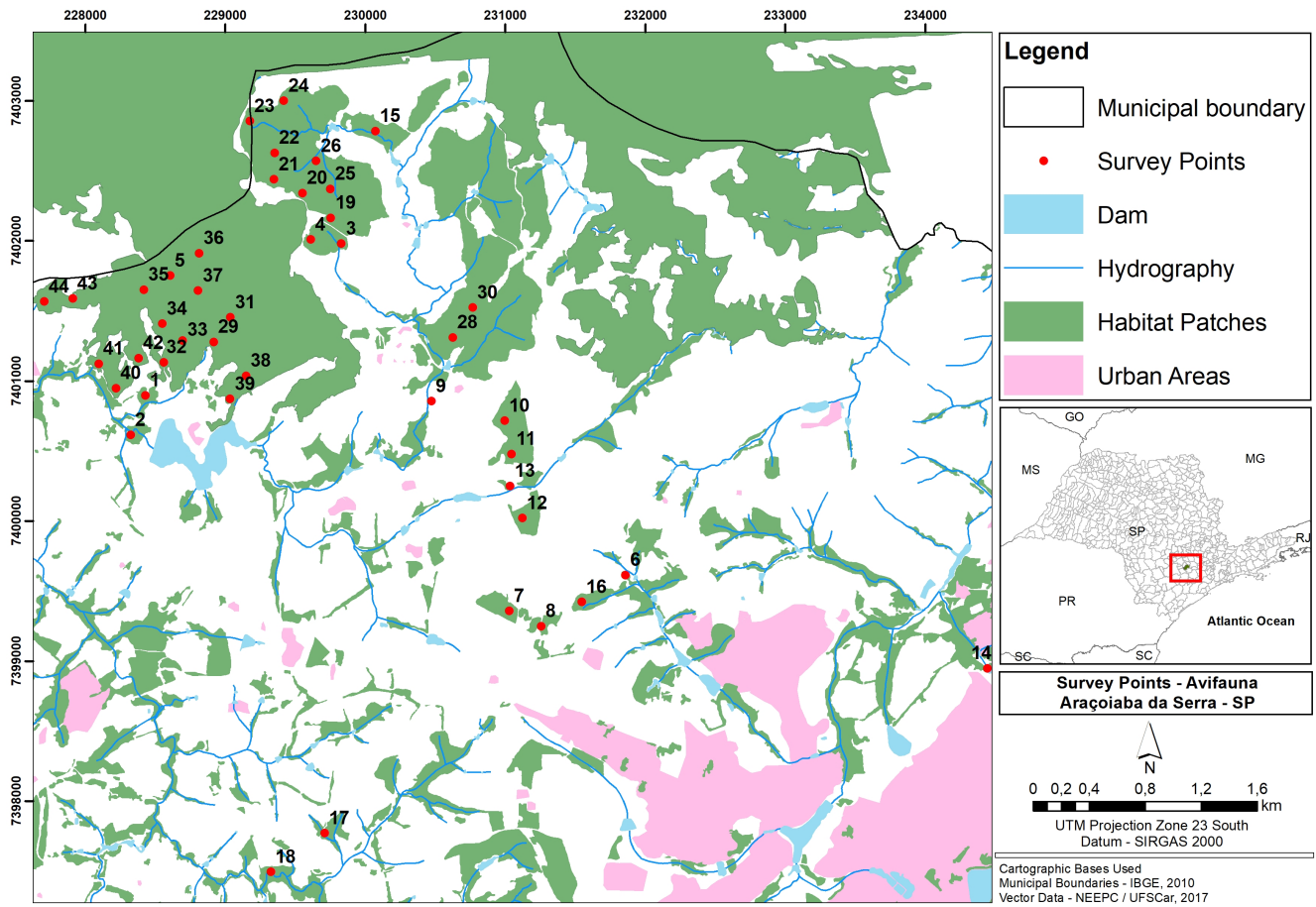


Figure 2. Fixed point counts in Araçoiaba da Serra, São Paulo, Brazil.

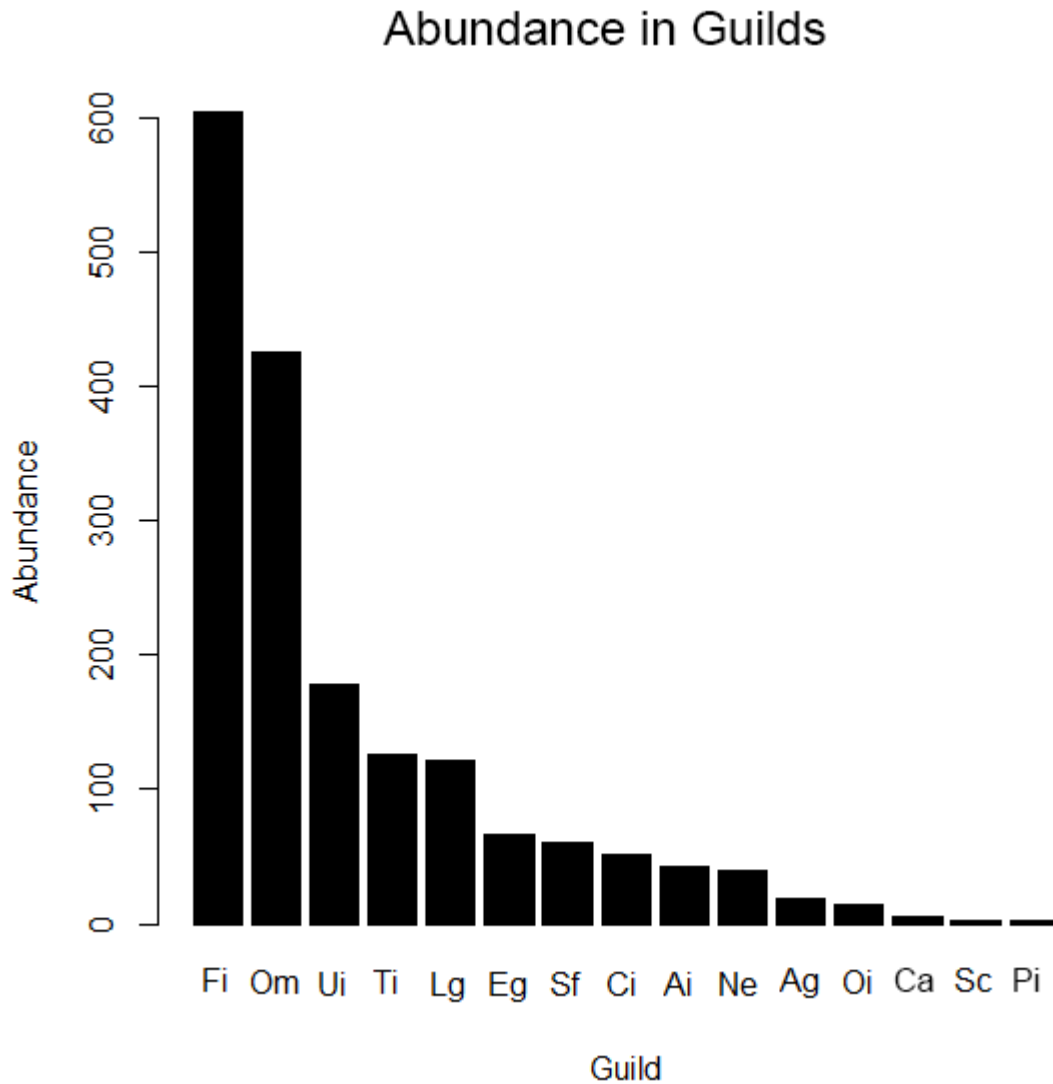


Figure 3. Abundance of birds in each guild foraging in a region of Atlantic Forest in southeastern Brazil. Legend: foliage insectivores (Fi); omnivores (Om); understory insectivores (Ui); trunk climbers insectivorous (Ti); large frugivores (Lg); edge granivores (Eg); aerial insectivorous (Ai); canopy insectivores (Ic); nectarivores (Ne); small frugivores (Sf); open area granivores (Ag); open area insectivores (Oi); scavengers (Sc); carnivores (Ca) and piscivores (Pi).



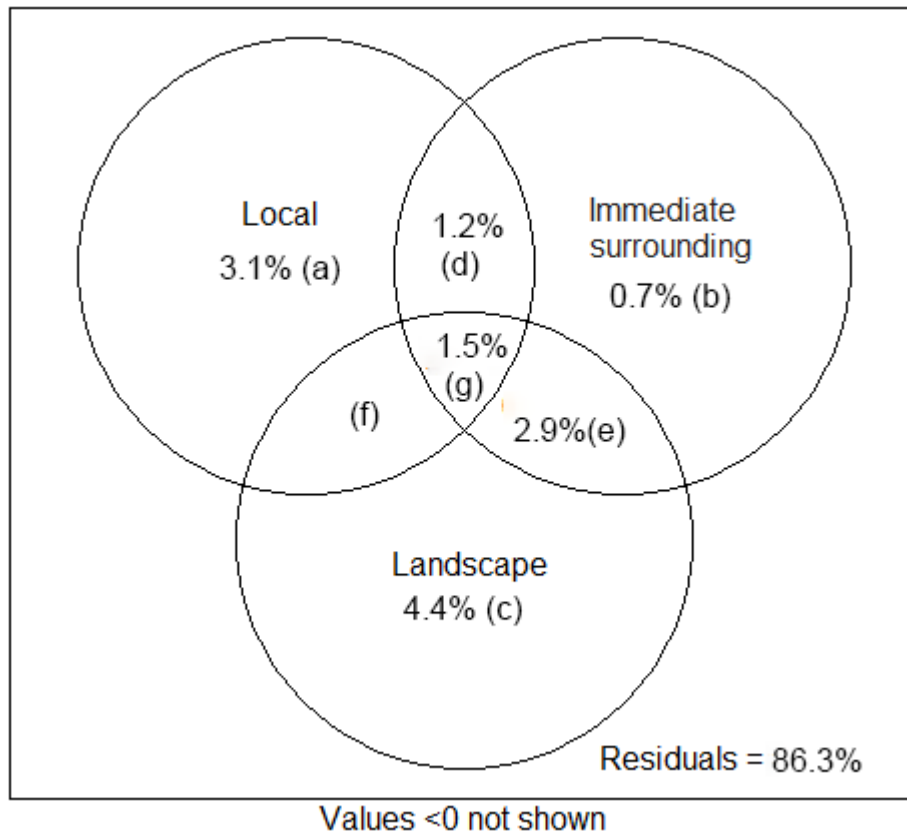
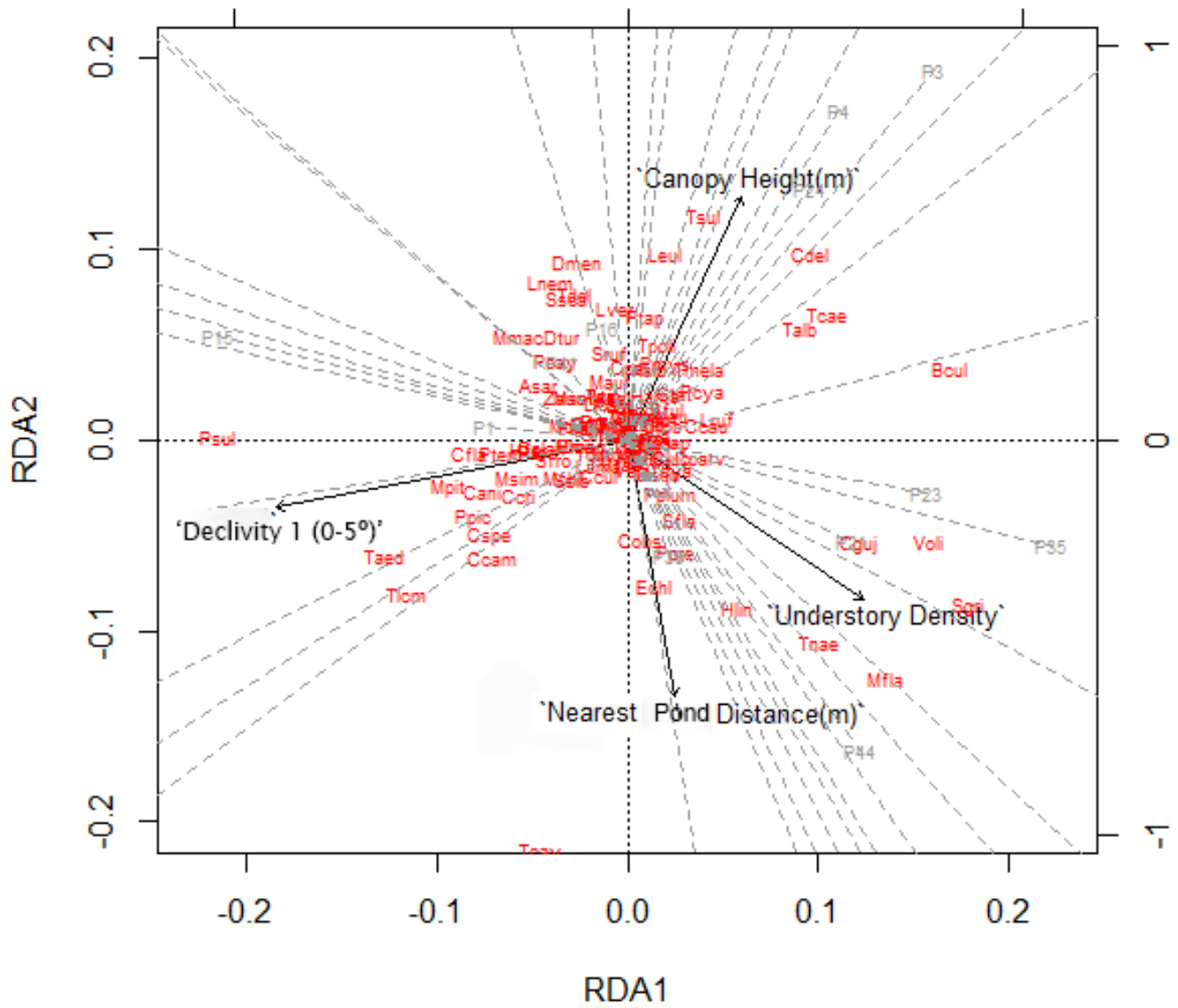


Figure 4. Venn diagram displaying the results of the variation partitioning analysis of the three explanatory scales on the abundance of birds. The explanatory variables were local (X1), landscape (X2) immediate surrounding and (X3) habitat features. The a, b, and c fractions represent the portion of bird abundances variation explained exclusively by local, immediate surrounding and landscape zone, respectively; the intersections indicate the fraction of variability in the response variable accounted for by both explanatory variables sets; and indicates the fraction of the variation that were not explained by the model.



.Figure 5. Ordination plot of the partial RDA analysis with the effects of the explanatory variables that best explain the variation of bird abundances in a region of the Atlantic Forest in south-eastern Brazil.



Table S1. Bird species recorded by fixed point counts (Jun-Sep 2017 and Oct 2017- Feb 2018).

Legend: Index of Punctual Abundance (IPA), foliage insectivores (Fi); omnivores (Om); understory insectivores (Ui); trunk climbers insectivorous (Ti); large frugivores (Lg); edge granivores (Eg); aerial insectivorous (Ai); canopy insectivores (Ic); nectarivores (Ne); small frugivores (Sf); open area granivores (Ag); open area insectivores (Oi); scavengers (Sc); carnivores (Ca) and piscivores (Pi).

Family	Specie	Acronym	IPA	Guild
Tinamidae	<i>Crypturellus parvirostris</i>	Cparv	0,159	Ag
	<i>Crypturellus tataupa</i>	Ctat	0,091	Fi
Cracidae	<i>Penelope obscura</i>	Pobs	0,114	Lg
Cathartidae	<i>Coragyps atratus</i>	Catr	0,068	Sc
Accipitridae	<i>Rupornis magnirostris</i>	Rmag	0,023	Ca
Rallidae	<i>Aramides saracura</i>	Asar	0,114	Om
Columbidae	<i>Columbina squammata</i>	Csqu	0,045	Ag
	<i>Patagioenas picazuro</i>	Ppic	2,25	Lg
	<i>Patagioenas cayennensis</i>	Pcaye	0,273	Lg
	<i>Zenaida auriculata</i>	Zaur	0,227	Ag
	<i>Leptotila verreauxi</i>	Lver	0,705	Eg
	<i>Leptotila rufaxilla</i>	Lruf	0,432	Eg
	<i>Geotrygon montana</i>	Gmon	0,045	Eg
Cuculidae	<i>Piaya cayana</i>	Pcay	0,318	Om
	<i>Crotophaga ani</i>	Cani	0,182	Oi
	<i>Tapera naevia</i>	Tnae	0,50	Ui
Trochilidae	<i>Phaethornis pretrei</i>	Ppre	0,182	Ne
	<i>Eupetomena macroura</i>	Emac	0,136	Ne
	<i>Thalurania glaucopis</i>	Tgla	0,114	Ne
	<i>Leucochloris albicollis</i>	Lalb	0,091	Ne
	<i>Chlorostilbon lucidus</i>	Cluc	0,023	Ne
	<i>Stephanoxis lalandi</i>	Slal	0,068	Ne
Alcedinidae	<i>Megaceryle torquata</i>	Mtor	0,023	Pi
Galbulidae	<i>Galbula ruficauda</i>	Gruf	0,114	Ai
Ramphastidae	<i>Ramphastos toco</i>	Rtoc	0,250	Lg
Picidae	<i>Picumnus temminckii</i>	Ptem	0,091	Ti
	<i>Melanerpes candidus</i>	Mcan	0,068	Ti
	<i>Veniliornis spilogaster</i>	Vspi	0,227	Ti
	<i>Colaptes campestris</i>	Ccam	0,114	Ti
	<i>Campephilus robustus</i>	Crob	0,205	Ti
	<i>Hylatomus lineatus</i>	Hlin	0,068	Ti
Falconidae	<i>Caracara plancus</i>	Cplan	0,159	Ca
	<i>Milvago chimachima</i>	Mchi	0,045	Ca
Thamnophilidae	<i>Mackenziaena severa</i>	Msev	0,114	Fi
	<i>Dysithamnus mentalis</i>	Dmen	1,523	Fi
	<i>Thamnophilus doliatus</i>	Tdol	0,455	Fi
	<i>Thamnophilus caerulescens</i>	Tcae	1,364	Fi
Conopophagidae	<i>Conopophaga lineata</i>	Clin	0,091	Fi

Family	Specie	Acronym	IPA	Guild	
Furnariidae	<i>Sittasomus griseicapillus</i>	Sgri	1,205	Ti	
	<i>Dendrocincla turdina</i>	Dtur	0,50	Ti	
	<i>Xiphocolaptes albicollis</i>	Xalb	0,068	Ti	
	<i>Automolus leucophthalmus</i>	Aleu	0,568	Fi	
	<i>Sclerurus scansor</i>	Ssca	0,205	Fi	
	<i>Lochmias nematura</i>	Lnem	0,250	Ui	
	<i>Synallaxis ruficapilla</i>	Sruf	0,273	Fi	
	<i>Synallaxis spixi</i>	Sspi	0,023	Fi	
	<i>Synallaxis frontalis</i>	Sfro	0,250	Fi	
	<i>Cranioleuca pallida</i>	Cpal	0,136	Ti	
	<i>Syndactyla rufosuperciliata</i>	Srufo	0,023	Ui	
Pipridae	<i>Chiroxiphia caudata</i>	Ccau	0,477	Sf	
Tyrannidae	<i>Platyrinchus mystaceus</i>	Pmys	0,091	Ui	
	<i>Leptopogon amaurocephalus</i>	Lama	0,432	Ai	
	<i>Corythopsis delalandi</i>	Cde	0,818	Ui	
	<i>Myiornis auricularis</i>	Maur	0,250	Ui	
	<i>Tolmomyias sulphurescens</i>	Tsul	0,795	Fi	
	<i>Todirostrum poliocephalum</i>	Tpoli	0,091	Ci	
	<i>Phyllomyias plumbeiceps</i>	Pplum	0,045	Ui	
	<i>Todirostrum cinereum</i>	Tcin	0,159	Fi	
	<i>Hirundinea ferruginea</i>	Hfer	0,068	Ai	
	<i>Camptostoma obsoletum</i>	Cobs	0,455	Fi	
	<i>Myiarchus swainsoni</i>	Mswa	0,068	Ci	
	<i>Myiarchus tuberculifer</i>	Mtub	0,045	Fi	
	<i>Myiarchus ferox</i>	Mfer	0,091	Ci	
	<i>Myiarchus tyrannulus</i>	Mtyr	0,091	Ci	
	<i>Capsiempis flaveola</i>	Cflav	0,182	Ui	
	<i>Sirystes sibilator</i>	Ssib	0,250	Ci	
	<i>Pitangus sulphuratus</i>	Psul	0,864	Om	
	<i>Myiodynastes maculatus</i>	Mmac	0,409	Om	
	<i>Megarynchus pitangua</i>	Mpit	0,273	Om	
	<i>Myiozetetes similis</i>	Msim	0,295	Om	
	<i>Tyrannus melancholicus</i>	Tmel	0,250	Ci	
	<i>Myiophobus fasciatus</i>	Mfas	0,159	Ui	
	<i>Fluvicola nengeta</i>	Fnen	0,136	Oi	
		<i>Lathrotriccus euleri</i>	Leul	0,523	Ui
	Vireonidae	<i>Cyclarhis gujanensis</i>	Cguj	2,068	Om
		<i>Hylophilus amaurocephalus</i>	Hama	0,091	Ci
<i>Vireo olivaceus</i>		Voli	1,023	Fi	
Corvidae	<i>Cyanocorax cristatellus</i>	Ccri	0,091	Om	
Cathartidae	<i>Pygochelidon cyanoleuca</i>	Pcya	0,295	Ai	
	<i>Progne tapera</i>	Ptap	0,159	Ai	
Troglodytidae	<i>Troglodytes aedon</i>	Taed	0,50	Fi	
Turdidae	<i>Turdus leucomelas</i>	Tlcm	2,364	Om	
	<i>Turdus albicollis</i>	Talb	0,386	Om	
	<i>Turdus rufiventris</i>	Truf	0,227	Om	
	<i>Turdus amaurochalinus</i>	Tama	0,205	Om	

<b>Family</b>	<b>Specie</b>	<b>Acronym</b>	<b>IPA</b>	<b>Guild</b>
Mimidae	<i>Mimus saturninus</i>	Msat	0,182	Om
Passerellidae	<i>Zonotrichia capensis</i>	Zcap	0,136	Eg
Parulidae	<i>Setophaga pitayumi</i>	Spit	0,409	Fi
	<i>Basileuterus culicivorus</i>	Bcul	4,068	Fi
	<i>Myiothlypis flaveola</i>	Mfla	1,091	Ui
	<i>Myiothlypis leucoblephara</i>	Mleu	0,409	Fi
Thraupidae	<i>Thraupis sayaca</i>	Tsay	0,295	Om
	<i>Thraupis cyanopectera</i>	Tcya	0,091	Om
	<i>Tangara palmarum</i>	Tpal	0,068	Om
	<i>Tangara cayana</i>	Tcay	1,114	Om
	<i>Nemosia pileata</i>	Npil	0,045	Ci
	<i>Conirostrum speciosum</i>	Cspe	0,182	Ci
	<i>Sicalis flaveola</i>	Sfla	0,114	Eg
	<i>Coryphospingus cucullatus</i>	Ccul	0,068	Eg
	<i>Tachyphonus coronatus</i>	Tcor	0,068	Om
	<i>Ramphocelus carbo</i>	Rcar	0,182	Om
	<i>Dacnis cayana</i>	Dcay	0,045	Sf
	<i>Saltator fuliginosus</i>	Sful	0,045	Om
	<i>Coereba flaveola</i>	Cfla	0,295	Ne
	<i>Trichothraupis melanops</i>	Tmela	0,205	Sf
	<i>Tersina viridis</i>	Tvir	0,159	Sf
	<i>Euphonia chlorotica</i>	Echl	0,477	Sf
<i>Habia rubica</i>	Hrub	0,114	Ui	